Mobile monitoring of personal NO\textsubscript{x} exposures during scripted daily activities in Chicago, IL

Jiayao Xu\textsuperscript{1}, Han Jiang\textsuperscript{1}, Haoran Zhao\textsuperscript{1}, Brent Stephens\textsuperscript{1*}

\textsuperscript{1} Department of Civil, Architectural, and Environmental Engineering, Illinois Institute of Technology, Chicago, IL USA

Abstract

Elevated ambient concentrations of nitrogen oxides (NO\textsubscript{x}), including nitric oxide (NO) and nitrogen dioxide (NO\textsubscript{2}), are associated with a wide range of adverse human health effects. Most studies have investigated these associations using ambient NO\textsubscript{2}/NO\textsubscript{x} measurements from fixed-site monitors or modeled ambient NO\textsubscript{2}/NO\textsubscript{x} concentrations. However, the majority of personal exposures to NO\textsubscript{2}/NO\textsubscript{x} occur in a variety of different microenvironments in which people spend most of their time. Previous studies have reported widely varying correlations between personal exposures and ambient NO\textsubscript{2}/NO\textsubscript{x} concentrations over various timescales. To add to the knowledge base of how personal NO/NO\textsubscript{2}/NO\textsubscript{x} exposures vary spatially, temporally, and within different microenvironments in an urban environment, we conducted roll-around mobile monitoring of NO/NO\textsubscript{2}/NO\textsubscript{x} with 1-minute resolution during 14 days of scripted activities in and around Chicago, IL. Activities involved time spent in three primary microenvironments: outdoors, indoors inside various building types, and in multiple modes of transportation including walking, personal vehicle, and public transit. Measurement were conducted at a higher time resolution than most prior microenvironmental monitoring studies using a recently developed direct UV absorbance NO/NO\textsubscript{2}/NO\textsubscript{x} monitor that is designed to minimize interferences that have been observed in some field campaigns using chemiluminescence monitors. The individual microenvironmental categories with the highest median NO\textsubscript{x} concentrations included four indoor environments and a variety of public transit environments. The individual transportation microenvironments with the highest median NO\textsubscript{x} concentrations were found aboard regional trains, largely driven by high NO from diesel locomotives. Correlations between microenvironmental NO/NO\textsubscript{2}/NO\textsubscript{x} measurements and simultaneous records from the nearest ambient monitor were extremely low, with coefficients of determination below 0.05 for each NO\textsubscript{x} constituent. These data further illustrate the limitations of relying on ambient site regulatory monitors to characterize personal NO/NO\textsubscript{2}/NO\textsubscript{x} exposures and provide further evidence that

\* Corresponding author. Tel: 1-312-567-3356
E-mail address: brent@iit.edu
personal monitoring is critical for accurately assessing personal exposure to NO$_x$.

**Keywords:** Nitrogen oxides; Human exposure; Personal exposure; Mobile samplers; Indoor air pollution
INTRODUCTION

Elevated ambient concentrations of nitrogen oxides (NO\textsubscript{x}), including nitric oxide (NO) and nitrogen dioxide (NO\textsubscript{2}), have been associated with a wide range of adverse human health effects including respiratory effects, cardiovascular effects, lung cancer, and mortality (U.S. EPA, 2016). Although most associations with adverse health effects have been made using measurements from fixed-site ambient NO\textsubscript{2}/NO\textsubscript{x} monitors or modeled ambient NO\textsubscript{2}/NO\textsubscript{x} concentrations, personal exposures to NO\textsubscript{2}/NO\textsubscript{x} are more complex, particularly in urban microenvironments that have a variety of NO\textsubscript{2}/NO\textsubscript{x} sources. First, because motor vehicle emissions are the single largest contributor to NO\textsubscript{2}/NO\textsubscript{x} concentrations in ambient air in the U.S. and traffic patterns are highly variable, there are typically high spatial and temporal gradients in ambient NO\textsubscript{2}/NO\textsubscript{x} concentrations that vary with the distance from central site monitors (Henderson et al., 2007; Novotny et al., 2011; Montagne et al., 2013). Second, people spend most of their time in microenvironments other than outdoors, including inside homes, offices, restaurants, and vehicles (Klepeis et al., 2001), all of which can have varying fractions of ambient NO\textsubscript{2}/NO\textsubscript{x} that infiltrates and persists (Dimitroulopoulou et al., 2001; Zota et al., 2005; Fabian et al., 2012). Third, there are many indoor sources of NO\textsubscript{2}/NO\textsubscript{x} in the various microenvironments in which people spend most of their time, including cooking and space-heating using natural gas and other fuels (Yang et al., 2004; Kornartit et al., 2010; Logue et al., 2014). The combination of these effects leads to indoor NO\textsubscript{2}/NO\textsubscript{x} exposures that are often higher than outdoors (Baxter, Clougherty, Laden, et al., 2007;
Baxter, Clougherty, Paciorek, et al., 2007) and personal exposures that are influenced by exposures in a number of different microenvironments (Lee et al., 2000). These issues also complicate our ability to perform accurate personal exposure assessments for ambient NO/NO$_2$/NO$_x$, particularly on a short-term basis.

Previous studies have reported widely varying correlations between personal or microenvironmental exposures and ambient NO$_2$/NO$_x$ concentrations, typically increasing with sampling duration. Some of these studies have shown moderate correlations between personal and ambient and/or indoor NO$_2$/NO$_x$ concentrations for some populations (Sørensen et al., 2005; Meng, Svendsgaard, et al., 2012), while others have shown almost no correlation (Quackenboss et al., 1986; Kousa et al., 2001; Lai et al., 2004; Meng, Williams, et al., 2012). Many previous personal or microenvironmental NO$_2$/NO$_x$ studies have been limited to long sampling intervals using passive integrated samplers and most have focused on either NO$_2$ or total NO$_x$ (Esplugues et al., 2010; Borge et al., 2016; Xu et al., 2016), which limits understanding of important spatiotemporal variations in personal exposures to NO/NO$_2$/NO$_x$ that could affect short-term health effects and elucidate contributions from various sources. Moreover, most field campaigns that have made microenvironmental NO/NO$_2$/NO$_x$ measurements with higher temporal resolution used chemiluminescence monitors, some of which have been shown to be subject to interference by species common to urban environments including HONO, HNO$_3$, and peroxyacyl nitrates.
Therefore, to add to the knowledge base of how personal NO/NO$_2$/NO$_x$ exposures vary spatially, temporally, and within different microenvironments in an urban environment, we conducted roll-around mobile monitoring of NO/NO$_2$/NO$_x$ with 1-minute resolution during 14 days of scripted activities in and around Chicago, IL. Measurements were made using a new direct UV absorbance NO/NO$_2$/NO$_x$ monitor that is designed to minimize interferences that have been observed in field campaigns using chemiluminescence monitors. Scripted activities were designed to capture time spent in three primary microenvironments: outdoors, indoors inside various building types, and in multiple modes of transportation including walking, personal vehicle, and various modes of public transit. Results are intended to more accurately demonstrate the spatiotemporal variability in personal NO$_x$ exposures encountered during typical daily activity in an urban environment and to improve knowledge of how personal NO/NO$_2$/NO$_x$ exposures correlate with ambient central-site monitors in urban environments.

**METHODS**

Measurements were made using a 2B Technologies Model 405 NO/NO$_2$/NO$_x$ direct UV absorbance analyzer installed horizontally inside a roll-around bag connected to a 12V lead-acid
car battery for mobile monitoring, similar to that described for mobile measurements of personal
ozone concentrations in Johnson et al. (2013) (Fig. 1). NO₂ is measured directly by the instrument
using absorbance at 405 nm, and NO is measured by alternative sequential conversion to NO₂
with internally generated O₃. Total NOₓ is calculated by adding the resulting NO and NO₂
concentrations. The instrument has a manufacturer reported limit of detection of ~1 ppb and an
accuracy of 2 ppb or 2% of the reading, whichever is greater. The instrument logged at 1-minute
intervals for all measurement periods.

A 1 m length of PTFE (Teflon™) tubing was used for the sample inlet, installed at a height of
~0.5 m off the ground, and Tygon™ tubing was used for the analyzer’s exhaust port (located on
the opposite side of the bag from the sampling inlet). The rechargeable battery allowed for
measurements for up to 10 hours on a full charge each day of monitoring. 2B recommends an
analyzer operating temperature range between 10 and 50°C. Therefore, the temperature and
relative humidity inside the case was monitored using an Onset HOBO U12 recording at 1-
minute intervals during all measurements. The minimum temperature was just above 10°C and
the maximum temperature was 32°C. Other instrument checks included daily zero checks prior to
measurements, with a new offset applied as necessary, as well as weekly NO span calibration
using a 2B Technologies Model 408 NO Calibration Source.
Two researchers conducted roll-around measurements during a total of 14 days spanning a period of approximately two months during winter and spring, March 2016 through May 2016. During 10 of the sampling days, measurements were made for approximately 10 consecutive hours following a variety of scripted activities in and around Chicago, IL. A shorter sampling period was used for the remaining four days, which involved only a few hours of sampling near the main campus of Illinois Institute of Technology in Chicago, IL. Scripted activities were designed to capture a wide variety of typical behaviors and microenvironments encountered by residents of Chicago, including travel via multiple modes of transportation (e.g., personal vehicle, city bus, subway, elevated train, regional commuter train, taxi, and walking), residential activities (e.g., inactive periods indoors and cooking activities), work/school activities (e.g., attending class, working in a laboratory, or working in an office building), and dining in restaurants (e.g., both fast food and sit-down). Each activity was scheduled to last at least 10 minutes to ensure adequate data collection in each microenvironment, although many activities were conducted for longer periods of time. A smartphone application (Lat Long) was used to record the latitude and longitude of each measurement location during sampling. Each day’s route is shown graphically in Fig. 2, including a few longer suburban routes, several routes between IIT’s main campus and downtown Chicago, and several routes within downtown Chicago.
Time-series data from the roll-around monitor were downloaded every day after sampling. Data processing involved labeling each time-series data point with the specific location, latitude, longitude, and time in which sampling took place, notes on the type of activities that were present in a particular microenvironment, and the corresponding temperature, relative humidity, and ambient NO/NO$_2$/NO$_x$ concentration data from the analyzer. Each location and activity was then coded with more generalized descriptions and individual locations were grouped into the following three primary and nine secondary microenvironmental (ME) categories listed in Table 1.

Measurements were also conducted immediately outside of one of two local regulatory monitors during the majority of test days, typically for about one hour. This co-location period served to provide a check on the comparability of the roll-around analyzer and the local federal regulatory monitors. Of the five regulatory ambient NO/NO$_2$/NO$_x$ monitors located within Cook County, two sites were used for comparison: an urban site at 321 S Franklin Street in downtown Chicago, IL (visited 11 times), and a suburban site at 750 Dundee Road in the suburb of Northbrook, IL near O’Hare airport (visited once) (IL EPA, 2014). Both locations utilize a Teledyne API Model T200 chemiluminescence NO/NO$_2$/NO$_x$ analyzer, which is designated by the U.S. EPA as an automated federal reference method (FRM), with a sample inlet height of ~6 m. Hourly-averaged data from these regulatory monitors were kindly provided by personnel at
the State of Illinois Environmental Protection Agency. We should note that these data have not yet been assessed for meeting quality assurance thresholds through the EPA’s annual data certification process.

RESULTS AND DISCUSSION

After data processing, there were a total of nearly 4000 1-minute average samples, providing approximately 65 hours of useful microenvironmental NO\textsubscript{x} concentration data for analysis.

Co-Location Comparisons to Ambient Regulatory Monitoring Stations

Fig. 3 shows the resulting hourly average concentrations measured concurrently with the roll-around NO/NO\textsubscript{2}/NO\textsubscript{x} analyzer and the two ambient regulatory monitoring stations. Results shown from the roll-around monitor are averages and standard deviations of the 1-minute data summarized over the hour that was spent immediately outside the regulatory monitoring stations. Data from the ambient regulatory monitoring station include either one data point spanning the same hour during which the sampling team was outside the station, or the average and standard deviation across two hourly data points spanning the hours that the team was present outside the station.

Correlations between the roll-around analyzer and the regulatory monitors differed for NO\textsubscript{2} and NO\textsubscript{x}, as might be expected due to several possible interferences noted in the NO\textsubscript{x} FRM
(ASTM, 2005), but surprisingly also for NO. Coefficients of determination (R² values) for NO, NO₂, and NOₓ were 0.29, 0.56, and 0.69, respectively. Moreover, NOₓ measurements with the roll-around analyzer were slightly lower than the regulatory monitor (slope = 0.82), much lower for NO₂ (slope = 0.51), and higher for NO (slope = 1.44). Discrepancies between the monitors are likely due to a combination of differences in measurement methods (e.g., chemiluminescence vs. UV), inlet sample heights (e.g., ~0.5 m vs. ~6 m), and sampling intervals (e.g., 1-minute vs. 1-hour) and sampling timeframes that did not exactly overlap that may have captured different temporal phenomena such as highly varying traffic sources.

**Detailed Microenvironmental Comparisons**

Fig. 4 shows an example of time-series NO and NO₂ data collected on March 22, 2016 from both the roll-around monitor (at 1-minute average intervals) and the nearest fixed-site ambient regulatory monitor (at 1-hour average intervals, measured at the 321 S Franklin Street monitor downtown). The roll-around data reflect measurements made in several different microenvironments, with the highest NO₂ peaks occurring in indoor microenvironments (including retail and educational settings), the highest combined NO/NO₂ peaks occurring in transportation microenvironments (chiefly public transit), and the highest NO peaks occurring in outdoor microenvironments. None of these peak values were reflected in the hourly average concentrations measured at the fixed-site monitor.
Fig. 5 shows distributions of 1-minute average NO, NO₂, and NOₓ concentrations measured in each of the 9 individual (i.e., secondary) microenvironmental categories listed in Table 1. These same distributional data are also summarized in Table 2 by the number of observations (i.e., 1-minute interval data points in each category) and summary statistics (i.e., mean, standard deviation, and 10th, 50th, and 90th percentiles).

The individual microenvironmental categories with the five highest median NOₓ concentrations included four indoor environments and a variety of public transit environments. The median 1-minute average NOₓ concentration was highest in the residential microenvironments, which included measurements made in a bedroom and a kitchen inside an apartment unit while occupants were cooking on a natural gas stove (median NOₓ = 97 ppb, mostly NO). The next highest median microenvironmental NOₓ concentrations were those in the public transit category (median NOₓ = 27 ppb, mostly NO), which included a combination of measurements in regional train cars, local elevated train cars, local underground subway train cars, outdoor elevated train platforms, and underground subway and regional rail stations.

Outdoor measurements had the fourth lowest median NOₓ value of 17 ppb. However, peak 1-minute average concentrations of NO/NO₂/NOₓ were all highest outdoors and in one of the retail environments visited, with single readings reaching as high as 200 ppb for NO and NO₂ and as high as 300 ppb for NOₓ.
Results (i.e., p-values) from non-parametric statistical comparisons of \( \text{NO, } \text{NO}_2, \text{and } \text{NO}_x \) concentrations measured in each of the 9 individual microenvironments made using two-sample Wilcoxon rank-sum (i.e., Mann-Whitney) tests are also shown in Table 3. The majority of comparisons revealed statistically significant differences in \( \text{NO, } \text{NO}_2, \text{and } \text{NO}_x \) concentrations between the individual microenvironments. The microenvironmental comparisons that did not yield statistically significant differences in at least one measure of \( \text{NO}/\text{NO}_2/\text{NO}_x \) were (i) commercial buildings, restaurants, and retail stores, and (ii) outdoors, educational buildings, parking garages, and personal vehicles.

**Summary of Microenvironmental Comparisons.**

Fig. 6 shows distributions of the same 1-minute average \( \text{NO, } \text{NO}_2, \text{and } \text{NO}_x \) data grouped by the three primary microenvironmental categories: transportation, indoor, and outdoor. Median \( \text{NO}_x \) concentrations were highest in the transportation microenvironments (median = 26 ppb), followed by the indoor environments (median = 21 ppb), and lowest in the outdoor environments (median = 17 ppb). Similar patterns were also observed for \( \text{NO} \), as \( \text{NO} \) drove most of the variability in total \( \text{NO}_x \). Differences in \( \text{NO} \) and \( \text{NO}_x \) between each microenvironment were all highly statistically significant (\( p <0.0001 \) according to a two-sample Wilcoxon rank-sum, i.e., Mann-Whitney, test). \( \text{NO}_2 \) distributions were more similar across each of the three microenvironmental categories, although differences in \( \text{NO}_2 \) between indoor and outdoor and
indoor and transportation microenvironments were highly statistically significant (p <0.0001 for both). Differences between outdoor and transportation microenvironments were not as highly statistically significant (p = 0.006). Peak 1-min values of NO/NO₂/NOₓ were quite similar across all microenvironmental categories, suggesting that NO, NO₂, and NOₓ concentrations in excess of 100, 50, and 150 ppb, respectively, can all be encountered at times in each type of microenvironment depending on the nearby source characteristics.

Looking more closely into the indoor and transportation microenvironments, Fig. 7 and Fig. 8 show distributions of 1-minute average NO, NO₂, and NOₓ concentrations measured in six specific categories of indoor microenvironments and eight specific categories of transportation microenvironments, respectively. Residential indoor microenvironments had the highest median NOₓ concentrations, while all other indoor microenvironments were similar to each other. Interestingly, the lowest indoor NOₓ concentrations were observed in the above ground, open air parking garages that were visited. The individual transportation microenvironments with the highest median NOₓ concentrations were surprisingly found aboard regional trains, largely driven by high NO from diesel locomotives. The transportation microenvironments with the next highest median NOₓ concentrations were underground train stations (e.g., subway and/or regional rail). Personal vehicles and outdoor train platforms had the lowest median NOₓ concentrations in this sample.
Correlations Between Microenvironmental Measurements and Ambient Regulatory Monitors

Finally, Fig. 9 shows correlations between hourly average records of NO, NO\textsubscript{2}, and NO\textsubscript{x} taken from the 321 S Franklin Street ambient regulatory monitor in downtown Chicago and the simultaneous microenvironmental measurements made during those same time periods (regardless of location). This ambient monitor was chosen for comparison because it is the nearest monitor for the vast majority of the microenvironmental measurements that were made (Fig. 2). Because the regulatory monitor only provides a single hourly average value, roll-around microenvironmental NO\textsubscript{x} data were matched to the EPA data on an hourly basis via matching time stamps, and an average and standard deviation were obtained using as many 1-minute interval data points as were available in each hour. Comparisons between the roll-around and simultaneous ambient monitor data in Fig. 9 were limited to those hours that had at least 30 data points (i.e., 30 minutes of 1-minute interval data recorded during the same hourly timestamp as that recorded for the EPA monitor). This provided a total of 72 simultaneously recorded hourly NO\textsubscript{x} concentrations for comparison.

Correlations between microenvironmental NO/NO\textsubscript{2}/NO\textsubscript{x} measurements and simultaneous records from the nearest ambient monitor were extremely low, with $R^2$ values below 0.05 for all comparisons. This is consistent with observations from several prior studies that have reported essentially no correlation between personal NO\textsubscript{2}/NO\textsubscript{x} exposures and simultaneous NO\textsubscript{2}/NO\textsubscript{x}
measurements from local ambient monitoring stations or nearby outdoor measurements (Quackenboss et al., 1986; Kousa et al., 2001; Lai et al., 2004; Meng, Williams, et al., 2012).

These data further illustrate the limitations of relying on ambient site regulatory monitors to characterize personal NO/NO$_2$/NO$_x$ exposures and provide further evidence that personal monitoring is critical for accurately assessing personal exposure.

**Limitations**

There are several limitations to this study, as well as limitations to applicability of the measurement methods used here. First, the roll-around monitoring system we used is portable, yet bulky enough that it is not easily carried from one place to another in some areas of typical urban environments (e.g., up and down stairs). Second, measurements are limited to only a short time frame of about two weeks worth of data collection and limited only to the specific locations in and around Chicago, IL. These data may not be representative for other urban environments. Third, the comparison between the roll-around monitor co-located near the fixed-site regulatory monitor cannot be taken as a direct side-by-side comparison because of differences in sampling inlet heights. Last, each microenvironment was sampled for a relatively short period of time to capture a wide variety of activities, so they may not be representative of longer-term exposures.

**CONCLUSIONS**
In this work, roll-around mobile monitoring of NO/NO\textsubscript{2}/NO\textsubscript{x} was conducted with 1-minute resolution during 14 days of scripted activities in and around Chicago, IL. Results demonstrated that residential exposures and exposures in certain types of transit (e.g., regional train and city bus) are likely to drive NO/NO\textsubscript{2}/NO\textsubscript{x} exposures during typical daily activities in and around Chicago, IL. Correlations between microenvironmental NO/NO\textsubscript{2}/NO\textsubscript{x} measurements and simultaneous records from the nearest ambient monitor were extremely low, which further illustrates the limitations of relying on ambient site regulatory monitors to characterize personal NO/NO\textsubscript{2}/NO\textsubscript{x} exposures and provide further evidence to a growing body of literature that personal monitoring is critical for accurately assessing personal exposure to NO\textsubscript{x}.

ACKNOWLEDGMENTS

We would like to thank Chris Price at the Illinois EPA for providing NO\textsubscript{x} concentration data from local ambient monitoring stations, Will Ollison from the American Petroleum Institute (API) for his helpful comments on a draft of this work, and API for their support of this work via an unrestricted gift in support of the research activities of Brent Stephens.

DISCLAIMER

Reference to any companies or specific commercial products does not constitute endorsement.

REFERENCES


Office of Research and Development, U.S. Environmental Protection Agency, Research Triangle Park, NC.


Table Captions

**Table 1.** List of three primary and nine secondary microenvironment (ME) categories in which measurements were made.

**Table 2.** Summary statistics of 1-min average NO, NO$_2$, and NO$_x$ concentrations measured in 9 different types of microenvironments.

**Table 3.** p-values resulting from non-parametric statistical comparisons of NO, NO$_2$, and NO$_x$ concentrations measured in each of the 9 individual microenvironments using two-sample Wilcoxon rank-sum (i.e., Mann-Whitney) tests.

<table>
<thead>
<tr>
<th>Primary ME</th>
<th>Secondary ME</th>
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</thead>
<tbody>
<tr>
<td>Indoors</td>
<td>Residential buildings</td>
</tr>
<tr>
<td></td>
<td>Commercial buildings</td>
</tr>
<tr>
<td></td>
<td>Retail buildings</td>
</tr>
<tr>
<td></td>
<td>Educational buildings</td>
</tr>
<tr>
<td></td>
<td>Restaurants</td>
</tr>
<tr>
<td></td>
<td>Parking garages (all above-ground)</td>
</tr>
<tr>
<td>Transportation</td>
<td>Personal vehicles</td>
</tr>
<tr>
<td></td>
<td>Public transit (e.g., on a regional or local train or entrance platform)</td>
</tr>
<tr>
<td>Outdoors</td>
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</tbody>
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Table 2. Summary statistics of 1-min average NO, NO$_2$, and NO$_x$ concentrations measured in 9 different types of microenvironments.

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<tr>
<th>Location type</th>
<th>N</th>
<th>Mean (SD)</th>
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<th>50$^{th}$ percentile</th>
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Table 3. P-values resulting from non-parametric statistical comparisons of NO, NO$_2$, and NO$_x$ concentrations measured in each of the 9 individual microenvironments using two-sample Wilcoxon rank-sum (i.e., Mann-Whitney) tests.

<table>
<thead>
<tr>
<th>Nitric oxide (NO)</th>
<th>Residential</th>
<th>Commercial</th>
<th>Restaurant</th>
<th>Retail</th>
<th>Outdoor</th>
<th>Educational</th>
<th>Parking garage</th>
<th>Personal vehicle</th>
<th>Public transit</th>
</tr>
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<tbody>
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<tr>
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<td>&lt;0.0001</td>
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<table>
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<th>Restaurant</th>
<th>Retail</th>
<th>Outdoor</th>
<th>Educational</th>
<th>Parking garage</th>
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<th>Public transit</th>
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<th>Retail</th>
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</table>


Figure Captions

Fig. 1. NOx analyzer installed in a portable, roll-around case.

Fig. 2. Map of each sampling day’s routes: a) zoomed out to include surrounding suburban trips; b) zoomed in to include only trips between IIT’s main campus and downtown Chicago, IL; and c) zoomed in to include only trips within downtown Chicago, IL.

Fig. 3. Hourly average concentrations resulting from co-location measurements alongside two ambient regulatory monitors in Cook County.

Fig. 4. Example time-series NOx data collected on March 22, 2016 from (a) the roll-around monitor (sampling interval = 1 minute) and (b) the nearest fixed-site regulatory monitor located at 321 S Franklin Street in Chicago (sampling interval = 1 hour). Microenvironmental categories are marked for several periods of roll-around sampling.

Fig. 5. Distributions of 1-min average NO, NO2, and NOx concentrations measured in 9 different types of microenvironments.

Fig. 6. Distributions of 1-min average NO, NO2, and NOx concentrations measured in 3 main categories of microenvironments (n = 1010 for transportation, n = 2125 for indoor, and n = 707 for outdoor).

Fig. 7. Distributions of 1-min average NO, NO2, and NOx concentrations measured in 6 specific categories of indoor microenvironments.

Fig. 8. Distributions of 1-min average NO, NO2, and NOx concentrations measured in 8 specific categories of transportation microenvironments.

Fig. 9. Correlations between hourly average (± SD) microenvironmental NO, NO2, and NOx measurements (minimum of 30 1-minute interval data points) and concurrent average hourly concentrations taken from the EPA ambient regulatory monitoring station at 321 S Franklin Street, downtown Chicago, IL (n = 72).
Fig. 1. NO₅ analyzer installed in a portable, roll-around case.
Fig. 2. Map of each sampling day’s routes: a) zoomed out to include trips both within downtown Chicago, IL and to surrounding suburban areas; b) zoomed in to include only trips between IIT’s main campus and downtown Chicago, IL; and c) zoomed in to include only trips within downtown Chicago, IL. Ambient regulatory monitor locations are marked as 750 Dundee Road and 321 S Franklin Street.
Fig. 3. Hourly average concentrations resulting from co-location measurements alongside two ambient regulatory monitors in Cook County.
Fig. 4. Example time-series NO\(_x\) data collected on March 22, 2016 from (a) the roll-around monitor (sampling interval = 1 minute) and (b) the nearest fixed-site regulatory monitor located at 321 S Franklin Street in Chicago (sampling interval = 1 hour). Microenvironmental categories are marked for several periods of roll-around sampling.
Fig. 5. Distributions of 1-min average NO, \( \text{NO}_2 \), and \( \text{NO}_x \) concentrations measured in 9 different types of microenvironments.
Fig. 6. Distributions of 1-min average NO, NO$_2$, and NO$_x$ concentrations measured in 3 main categories of microenvironments (n = 1010 for transportation, n = 2125 for indoor, and n = 707 for outdoor).
Fig. 7. Distributions of 1-min average NO, NO₂, and NOₓ concentrations measured in 6 specific categories of indoor microenvironments.
Fig. 8. Distributions of 1-min average NO, NO$_2$, and NO$_x$ concentrations measured in 8 specific categories of transportation microenvironments.
Fig. 9. Correlations between hourly average (± SD) microenvironmental NO, NO$_2$, and NO$_x$ measurements (minimum of 30 1-minute interval data points) and concurrent hourly average concentrations taken from the EPA ambient regulatory monitoring station at 321 S Franklin Street, downtown Chicago, IL (n = 72).